Predicting resilience through the lens of competing adjustments to vegetation function

Manon E. B. Sabot1,2 | Martin G. De Kauwe1,2,3 | Andy J. Pitman1,2 | David S. Ellsworth4 | Belinda E. Medlyn4 | Silvia Caldararu5 | Sönke Zaehle5,6 | Kristine Y. Crous4 | Teresa E. Gimeno7,8 | Agnieszka Wujeska-Klause4,9 | Mengyuan Mu1,2 | Jinyan Yang4

1ARC Centre of Excellence for Climate Extremes, Sydney, New South Wales, Australia
2Climate Change Research Centre, University of New South Wales, Sydney, New South Wales, Australia
3School of Biological Sciences, University of Bristol, Bristol, UK
4Hawkesbury Institute for the Environment, Western Sydney University, Penrith, New South Wales, Australia
5Max Planck Institute for Biogeochemistry, Jena, Germany
6Michael Stifel Center Jena for Data-driven and Simulation Science, Jena, Germany
7CREAF, 08193 Bellaterra (Cerdanyola del Vallés), Catalonia, Spain
8Basque Centre for Climate Change (BC3), Leioa, Spain
9Urban Studies, School of Social Sciences, Penrith, New South Wales, Australia

Correspondence
Manon Sabot, Climate Change Research Centre, University of New South Wales, Sydney, NSW 2052, Australia.
Email: m.e.b.sabot@gmail.com

Abstract
There is a pressing need to better understand ecosystem resilience to droughts and heatwaves. Eco-evolutionary optimization approaches have been proposed as means to build this understanding in land surface models and improve their predictive capability, but competing approaches are yet to be tested together. Here, we coupled approaches that optimize canopy gas exchange and leaf nitrogen investment, respectively, extending both approaches to account for hydraulic impairment. We assessed model predictions using observations from a native Eucalyptus woodland that experienced repeated droughts and heatwaves between 2013 and 2020, whilst exposed to an elevated [CO2] treatment. Our combined approaches improved predictions of transpiration and enhanced the simulated magnitude of the CO2 fertilization effect on gross primary productivity. The competing approaches also worked consistently along axes of change in soil moisture, leaf area, and [CO2]. Despite predictions of a significant percentage loss of hydraulic conductivity due to embolism (PLC) in 2013, 2014, 2016, and 2017 (99th percentile PLC > 45%), simulated hydraulic legacy effects were small and short-lived (2 months). Our analysis suggests that leaf shedding and/or suppressed foliage growth formed a strategy to mitigate drought risk. Accounting for foliage responses to water availability has the potential to improve model predictions of ecosystem resilience.

KEYWORDS
drought, elevated CO2, gas exchange, hydraulic legacies, land surface models, leaf area index, nitrogen, optimization, plant optimality, vegetation models
1 | INTRODUCTION

Climate projections suggest trees will soon encounter unprecedented extreme hydro-climatic conditions, which opens up important questions about their resilience. The ability of trees to maintain and recover function in response to disturbance preconditions their chances of avoiding dieback, die-off, and ultimately, local extinction. Changes in the intensity, frequency, and timing of climate extremes (Masson-Delmotte et al., 2021) are already putting tree species at risk (Brodribb et al., 2020; McDowell et al., 2020). Droughts, for instance, have been implicated as one of the main drivers of an increasing trend in global forest mortality (Allen et al., 2015), and more widespread, frequent, and severe drought episodes are expected in the future (Ault, 2020; Masson-Delmotte et al., 2021; Ukkola et al., 2020). To gain insight into the future resilience of tree species, we need models capable of predicting functional adjustments as ecosystems undergo global change. In that context, process-based optimization approaches that shave away empirical assumptions and reduce reliance on detailed parameterization are gaining traction, as they could increase the predictive power of global vegetation models (Fisher & Koven, 2020; Harrison et al., 2021).

Plant optimality theory relies on the evolutionary assumption that vegetation function and structure adjust to maximize return on resource investment in relation to external drivers (Bloom, 1986; Givnish, 1986; Mäkelä et al., 2002). It provides a pathway for the development of new, integrated mechanistic theory, and thereby improved models. Optimality principles have demonstrated an ability to predict broad vegetation patterns within and across ecosystems (e.g., Caylor et al., 2004; Fyllas et al., 2014; Jucker et al., 2015; Langan et al., 2017). Efforts to date have largely borrowed from economics (Franklin et al., 2020; Wright et al., 2004), applied with a focus on traits and processes that determine average and/or long-term vegetation behaviours, e.g., average photosynthetic capacities (Maire et al., 2012; Smith et al., 2019), canopy height (Falster & Westoby, 2003; Larjavaara & Muller-Landau, 2012), and rooting depth (McMurtrie et al., 2012; Yang et al., 2016).

In land surface models (LSMs), two recently adopted optimization approaches are advancing the representation of dynamic ecosystem fluxes. First, theory around the optimal regulation of leaf nitrogen investment has been used to capture temporal changes in photosynthetic capacity, improving net/gross primary productivity predictions (Caldararu et al., 2020; Haverd et al., 2018). Second, substantial improvements in simulated canopy gas exchange have been achieved by optimally balancing photosynthetic gains and the hydraulic costs of transporting water between the roots and the leaves (Eller et al., 2020; Sabot et al., 2020). However, as much as optimality principles hold promise for future LSM developments, they are not completely exempt from shortcomings. Chiefly, the choice of fitness proxies over which to optimize function and the selection of relevant optimization timescales are critical to model outcomes.

In reality, it is unlikely plants can optimize all resource acquisition and develop strategies to cope with stress factors simultaneously (Stearns, 1989). Optimality principles that are successful in isolation may therefore not be successful when combined with competing principles that operate on different timescales, or when rendered “suboptimal” by environmental disturbances. The few studies that have considered multiple functional adjustments together have typically relied on a unique or dominant overarching optimization objective, such as to maximize long-term carbon profits (see Potkay et al., 2021; Schymanski et al., 2009, 2015). For example, although long-term optimization objectives may be suitable for predictions of decadal carbon and/or water balance, they might neglect nonlinear interactions between processes, leading to unrealistic short-term predictions (see Kerkhoff et al., 2004). To our knowledge, there has not been an attempt to examine whether competing optimality principles can operate interactively and consistently across timescales of application, nor when system legacy is involved (i.e., lingering effects from repeated disturbance).

Multiple evidence streams—satellites (Wigneron et al., 2020), eddy covariance (Schwalm et al., 2017), tree-rings (Peltier et al., 2021; Vanoni et al., 2016), and manipulation experiments (Zweifel et al., 2020)—have shown that drought impacts on forest carbon stocks and tree growth can last several years after droughts have concluded. One of the mechanisms most often invoked to explain these legacies is hydraulic damage. Hydraulic damage is sustained from incomplete and/or lagged recovery of xylem function when cavitation is significant in trees (Klein et al., 2018). Hydraulic damage can cause leaf drop and/or hinder foliage growth, and lead to canopy dieback or even tree mortality in severe cases (McDowell et al., 2008). Critically, hydraulic damage continues to impede water flow between the soil and the canopy after soil water is recovered. Integrating this type of system legacy with optimality approaches is a challenge we must tackle to ensure our models capture future climate risk to the vegetation appropriately.

In this study, we combine nitrogen-based and hydraulics-based plant optimality theories within a single LSM framework (Sabot et al., 2020), while also incorporating a representation of legacy effects from hydraulic damage. Doing so allows us to investigate interactions between competing optimization theories affected by disturbance, at the forefront of LSM development.

We use observations from the Eucalyptus Free Air CO2 Enrichment (EucFACE) experiment to parameterize and assess our model. The EucFACE experiment is set in a native, mature Eucalyptus woodland that is nutrient-limited and that strongly responds to water availability. For example, Duursma et al. (2016) demonstrated that leaf flushing events occurred shortly after significant rainfall, whereas the site leaf area index (LAI) declined steadily during periods of low rainfall. The period considered here (2013–2020) encompasses frequent droughts and heatwaves, including a multi-year drought from mid-2017 onwards.

Recurring droughts and heatwaves may not only affect the hydraulic transport system, but also leaf nitrogen investments in the photosynthetic apparatus and hence photosynthetic capacity (Damour et al., 2008; Hikosaka et al., 2006; Vaz et al., 2010). Theory and experimental evidence also suggest that, besides inducing some leaf nitrogen reallocation (Ainsworth & Rogers, 2007), increasing
atmospheric [CO₂] could alleviate plant hydraulic stress during drought (De Kauwe et al., 2021; Jiang et al., 2021). However, we do not know whether competing optimality principles can successfully predict multiple axes of functional change in response to change in [CO₂]. Observations of LAI, water fluxes, leaf water potential, leaf photosynthetic capacity, and leaf nitrogen content, available under both ambient and elevated [CO₂] at EucFACE, provide us with an opportunity to probe competing combinations of optimality principles (see above) and to thoroughly interrogate physiological responses to weather and climate extremes (i.e., drought and heatwaves).

Our specific objectives are:

(i) to infer the timescales over which vegetation function adjusts to maximize resource investment and/or resilience;
(ii) to test whether contrasting optimization approaches can be applied together, including when hydraulic legacies from climate extremes affect optimization outcomes;
(iii) to explore the relationship between drought-driven foliage dynamics and plant hydraulic function; and
(iv) to understand whether (i), (ii), and (iii) are affected by elevated [CO₂].

2 | MATERIALS AND METHODS

2.1 | Site description

EucFACE is located in a temperate-subtropical woodland of western Sydney, Australia (33°36′59″S, 150°44′17″E). The experiment consists of six 25-m diameter circular plots, hereafter referred to as “rings.” R2, R3, and R6 are exposed to ambient [CO₂] and act as controls (aCO₂ rings; c. 408 µmol mol⁻¹ for the study period), whereas R1, R4, and R5 are fumigated with CO₂ (eCO₂ rings; aCO₂ +150 µmol mol⁻¹). The canopy is 18–23 m tall and is dominated by Eucalyptus tereticornis. Tree growth is limited by low water and low nutrient availability, particularly phosphorus (Crous et al., 2015; Ellsworth et al., 2017). For a detailed site overview, see Gimeno et al. (2016) and Jiang et al. (2020b).

2.2 | Overview of the model

The model described in Sabot et al. (2020) was adapted from the Community Atmosphere Biosphere Land Exchange LSM (CABLEv2.0; De Kauwe et al., 2015). In brief, our model accounts for: canopy interception; radiation scattering into sunlit and shaded components of a single-layer canopy; within-canopy micrometeorology; optimal canopy gas exchange; plant hydraulic processes; soil evaporation; and soil hydrology over six discrete layers. In this study, the canopy interception (Methods S1) and soil hydrology (Methods S2) schemes are revised to better reflect observations at EucFACE (Gimeno et al., 2018; see below). We also extend the model by embedding a scheme that optimizes leaf nitrogen allocation, thus varying photosynthetic capacity, and by incorporating hydraulic legacies alongside our existing canopy gas exchange optimization scheme (see below).

2.2.1 | Canopy gas exchange

Gas exchange is simulated via an instantaneous stomatal optimization approach (Sperry et al., 2017) adapted for LSMs (Sabot et al., 2020). Sunlit and shaded components of the canopy are each assumed to optimally balance relative photosynthetic gains and relative hydraulic costs through their stomates:

\[
\max \left( \frac{A_n}{A_{n,max}} - \frac{k_{c,max} - k_{crit}}{k_{c,max} - k_{crit}} \right)
\]

where \(A_n\) (µmol m⁻² s⁻¹) is the net rate of carbon assimilation; \(A_{n,max}\) (µmol m⁻² s⁻¹) is the instantaneous maximum \(A_n\); there could be considering soil moisture limitations; \(k_{c,max}\) (mmol m⁻² s⁻¹ MPa⁻¹) is the maximum root-to-leaf hydraulic conductance attenuated by soil moisture stress at a specific instant; \(k_{crit}\) (mmol m⁻² s⁻¹ MPa⁻¹) is hydraulic conductance evaluated at a given leaf water potential \(\Psi_l\) (MPa); and \(k_{crit}\) (mmol m⁻² s⁻¹ MPa⁻¹) is the critically low threshold of hydraulic conductance that indicates xylem failure.

Except \(k_{crit}\) (constant), all the elements that makeup Equation (1) vary on an instantaneous basis. As the elements in Equation (1) are computed from system of equations that connects photosynthetic, stomatal, and plant hydraulic processes (Methods S3 and S4), all these processes are also solved for when Equation (1) is solved for.

2.2.2 | Leaf nitrogen allocation

We draw on work that links nitrogen stoichiometry to the coordination of photosynthetic capacities (Chen et al., 1993; Field, 1983; Maire et al., 2012; Medlyn, 1996) to: (i) optimize the amount of leaf nitrogen involved in photosynthesis (\(N_p\); mol m⁻²) and its distribution to photosynthetic compounds; whilst also (ii) coordinating key photosynthetic parameters, that is, \(V_{cmax}\) (µmol m⁻² s⁻¹), the maximum carboxylation rate \(V_{cmax}\) at 25°C, \(J_{max}\) (µmol m⁻² s⁻¹), the maximum rate of electron transport \(J_{max}\) at 25°C, and \(R_{d25}\) (µmol m⁻² s⁻¹), the day leaf respiration \(R_d\) at 25°C. \(N_p\) is divided into thylakoid nitrogen and soluble protein nitrogen compounds that link to \(V_{cmax}\) and \(J_{max}\) (Medlyn 1996; Methods S5), so our approach allows \(N_p\) and the photosynthetic parameters to co-vary with environmental conditions, even in the absence of an explicit nitrogen cycle.

Our model seeks to satisfy:

\[
\min (|A_l (|\Sigma_{ante} N_p|) - A_l (|\Sigma_{ante} N_p|))
\]

subject to

\[
A_n (|\Sigma_{ante} N_p|) > A_n (|\Sigma_{ante} N_p|_{-1})
\]

\[
0 < A_n < 0.5 V_{cmax}
\]
where $A_c$ (µmol m$^{-2}$ s$^{-1}$) and $A_l$ (µmol m$^{-2}$ s$^{-1}$) are the ribulose-1,5-bisphosphate (RuBP) carboxylation (Rubisco) limited and the RuBP regeneration limited rates of carbon assimilation obtained from the photosynthesis model (Methods S3); and $\Sigma_{ante}$ is a set of average antecedent leaf-to-canopy scalers and canopy light environment, canopy leaf temperature ($T_l$; °C), $\text{CO}_2$ in the leaf intercellular air spaces ($C_i$; Pa), and $R_d$ recorded daily under maximum sunlight.

Both $\Sigma_{ante}$ and $N_p$ vary over a chosen time period (see Section 2.3). Importantly, the $N_p$ and photosynthetic parameters optimized through Equation (2) do not impose model co-limitation through time because "future" environmental conditions will differ from $\Sigma_{ante}$. Equation (2) is bound by the inequalities given in Equations (3) and (4). Equation (3) signifies that the new optimized $N_p$ and photosynthetic parameters must a posteriori increase $A_n$ over the conditions given by $\Sigma_{ante}$, compared to the previous time period $N_p$ ($N_{p_{t-1}}$). Equation (4) avoids optimization "overshoots" in the ratio of $\dot{A}_{max25} : \dot{V}_{max25}$ at elevated $\text{CO}_2$ (see Medlyn, 1996) by approximating a triose phosphate utilization (TPU) limitation on the objective function (Collatz et al., 1991). Note, however, that no TPU limitation is applied in the photosynthesis model (Methods S3).

### 2.3 | Model simulations

#### 2.3.1 | Environmental drivers and model parameters

Our model was forced with in situ half-hourly meteorological data (photosynthetic photon flux density (PPFD), air temperature ($T_a$), vapour pressure deficit ($D_v$), precipitation, atmospheric pressure, and wind speed) and atmospheric $\text{CO}_2$ ($C_a$) that were recorded, gap-filled (<0.8% of raw data), and aggregated to 30-min timespans as per Yang et al. (2020) and Mu et al. (2021). Canopy phenology was prescribed by the observed leaf area index (LAI) at each ring. LAI was calculated following Yang et al. (2020) and Mu et al. (2021), by removing the woody contribution (0.8 m$^2$ m$^{-2}$) from the plant area index based on diffuse canopy transmittance and monthly leaf litter production (Duursma et al., 2016). Figure 51 shows major model drivers.

The model was parameterized as shown in Table 1; parameter values were either retrieved from the literature or estimated from field observations (Methods S6). Other model parameters that were kept to their default values are shown in Tables S1 and S2.

#### 2.3.2 | Timescales of adjustments to function

We ran 36 simulation experiments to identify the timescales over which competing adjustments to vegetation function interact most consistently. Our default model configuration assumes the optimization of canopy gas exchange only, with no change in leaf nitrogen and photosynthetic parameters over time and no hydraulic legacies. Other configurations optimize canopy gas exchange, plus leaf nitrogen allocation and/or hydraulic legacies (Figure 1).

The timescales relevant to optimizing canopy gas exchange are not examined, as previous work indicated less than 3-hourly intervals were appropriate (Sabot et al., 2020), and half-hourly intervals are therefore used. Five model configurations consider the optimal regulation of leaf nitrogen ($N_{opt}$) every 7th, 14th, 21st, 28th, or 35th day. Another five configurations assess hydraulic legacies ($H_{lag}$) every 14th, 30th, 60th, 90th, or 180th day. The remaining 25 configurations enable joint variations in leaf nitrogen and hydraulic legacies, by combining the timescales previously listed for each process. The following matrix summarizes the model configurations, that is, the processes accounted for in addition to the optimization of canopy gas exchange and their associated timescales of application:

$$\begin{bmatrix}
- & N_{opt7} & \cdots & N_{opt35} \\
H_{lag14} & H_{lag14}N_{opt7} & \cdots & H_{lag14}N_{opt35} \\
\vdots & \vdots & \ddots & \vdots \\
H_{lag180} & H_{lag180}N_{opt7} & \cdots & H_{lag180}N_{opt35}
\end{bmatrix}$$

where - denotes the default model configuration, and $N_{opt7}$ and $H_{lag14}$ stand for $N_{opt}$ every 7th day (or $N_{opt} = 7$ days) and $H_{lag}$ every 14th day (or $N_{lag} = 14$ days), respectively.
<table>
<thead>
<tr>
<th>Param.</th>
<th>Definitions</th>
<th>Units</th>
<th>Values*</th>
<th>aCO2</th>
<th>eCO2</th>
</tr>
</thead>
<tbody>
<tr>
<td>$\nu_{\text{max}25t_0}$</td>
<td>Maximum carboxylation rate, $V_{\text{cmax}}$, at 25°C</td>
<td>µmol C m⁻² s⁻¹</td>
<td>R2</td>
<td>92.7</td>
<td>92.1</td>
</tr>
<tr>
<td>$J_{\text{max}25t_0}$</td>
<td>Maximum rate of electron transport, $J_{\text{max}}$, at 25°C</td>
<td>µmol electron m⁻² s⁻¹</td>
<td>R2</td>
<td>152.1</td>
<td>138.9</td>
</tr>
<tr>
<td>$R_{425t_0}$</td>
<td>Day respiration rate, $R_d$, at 25°C</td>
<td>µmol C m⁻² s⁻¹</td>
<td>R2</td>
<td>1.4</td>
<td>1.1</td>
</tr>
<tr>
<td>$E_a$</td>
<td>Activation energy</td>
<td>J mol⁻¹</td>
<td>R2</td>
<td>66.386 (ν_{max}), 32292 (J_{max})</td>
<td></td>
</tr>
<tr>
<td>$\delta_S$</td>
<td>Entropy factor</td>
<td>J mol⁻¹ K⁻¹</td>
<td>R2</td>
<td>639.60 (ν_{max}), 638.06 (J_{max})</td>
<td></td>
</tr>
<tr>
<td>$\alpha_l$</td>
<td>Short wave (visible) leaf reflectivity</td>
<td>–</td>
<td>R2</td>
<td>0.125⁵</td>
<td></td>
</tr>
<tr>
<td>$c$</td>
<td>Curvature of the electron transport rate to light</td>
<td>–</td>
<td>R2</td>
<td>0.48⁵</td>
<td></td>
</tr>
<tr>
<td>$\eta_{\text{cat}}$</td>
<td>Specific activity of Rubisco</td>
<td>mol C mol⁻¹ Rubisco s⁻¹</td>
<td>R2</td>
<td>22.2⁶</td>
<td>26.6⁶</td>
</tr>
<tr>
<td>$P_x$</td>
<td>Water potential drop at x% xylem conductivity loss</td>
<td>MPa</td>
<td>R2</td>
<td>-4.07 (P_{50}); -5.50 (P_{50})</td>
<td></td>
</tr>
<tr>
<td>$k_{\text{max}}$</td>
<td>Maximum root-to-leaf hydraulic conductance</td>
<td>mmol m⁻² s⁻¹ MPa⁻¹</td>
<td>R2</td>
<td>1.02</td>
<td>1.37</td>
</tr>
<tr>
<td>$k_{\text{crit}}$</td>
<td>Hydraulic conductance indicative of xylem failure</td>
<td>mmol m⁻² s⁻¹ MPa⁻¹</td>
<td>R2</td>
<td>0.05$k_{\text{max}}$</td>
<td></td>
</tr>
<tr>
<td>$r_k$</td>
<td>Ratio of xylem recovery from embolism</td>
<td>–</td>
<td>R2</td>
<td>0.92</td>
<td>0.88</td>
</tr>
<tr>
<td>$z_{\text{top}}$</td>
<td>Surface soil layer depth</td>
<td>cm</td>
<td>R2</td>
<td>5</td>
<td></td>
</tr>
<tr>
<td>$z_{\text{eff,r}}$</td>
<td>Effective rooting depth</td>
<td>cm</td>
<td>R2</td>
<td>48.1</td>
<td>50.8</td>
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<tr>
<td>$\kappa_r$</td>
<td>Extinction coefficient for the root distribution</td>
<td>1 cm⁻¹</td>
<td>R2</td>
<td>0.94</td>
<td></td>
</tr>
<tr>
<td>$b_{\text{ch}}$</td>
<td>Clapp-Hornberger pore size distribution index</td>
<td>–</td>
<td>R2</td>
<td>5.13</td>
<td>4.06</td>
</tr>
<tr>
<td>$\Psi_{\text{dep}}$</td>
<td>Air entry point soil water potential</td>
<td>kPa</td>
<td>R2</td>
<td>-1.29</td>
<td>-1.32</td>
</tr>
<tr>
<td>$K_{\text{sat}}$</td>
<td>Soil hydraulic conductivity at saturation</td>
<td>µm s⁻¹</td>
<td>R2</td>
<td>16.5</td>
<td>18.0</td>
</tr>
<tr>
<td>$\Theta_{\text{sat}}$</td>
<td>Volumetric soil water content at saturation</td>
<td>m³ m⁻³</td>
<td>R2</td>
<td>0.39</td>
<td>0.38</td>
</tr>
<tr>
<td>$\Theta_{\text{fc}}$</td>
<td>Volumetric soil water content at field capacity</td>
<td>m³ m⁻³</td>
<td>R2</td>
<td>0.20</td>
<td>0.17</td>
</tr>
<tr>
<td>$\Theta_{\text{wtp}}$</td>
<td>Volumetric soil water content at the soil wilting point</td>
<td>m³ m⁻³</td>
<td>R2</td>
<td>0.09</td>
<td>0.06</td>
</tr>
<tr>
<td>$r_{\text{fl}}$</td>
<td>Fractional resistance to soil evaporation</td>
<td>–</td>
<td>R2</td>
<td>0.82</td>
<td>0.82</td>
</tr>
<tr>
<td>$R_{\text{canat}}$</td>
<td>Minimum rainfall for canopy saturation</td>
<td>mm ½h⁻¹</td>
<td>R2</td>
<td>0.12</td>
<td>0.21</td>
</tr>
<tr>
<td>$\kappa_{\text{ci}}$</td>
<td>Extinction coefficient of canopy interception</td>
<td>–</td>
<td>R2</td>
<td>0.70</td>
<td>0.59</td>
</tr>
</tbody>
</table>

Note: The values are informed by site observations for all the parameters except $\eta_{\text{cat}}$ and $k_{\text{crit}}$.

*Values are given by Yang et al. (2020) for EucFACE. $\alpha_l$ matches $a_{\text{abs}} = 1 - (\alpha_l + \tau)$ where $a_{\text{abs}}$ is the absorptance of photosynthetic photon flux density (Yang et al., 2020) and $\tau$ is leaf transmittivity (0.05 in our model).

*Values estimated from substrate-saturated carboxylation rates, Rubisco site counts, and totals of leaf nitrogen allocated to Rubisco reported by Sharwood et al. (2017) for upper canopy leaves of *Eucalyptus globulus*.

*Values derived by Peters et al. (2021) from hydraulic vulnerability curves constructed by benchtop dehydration for *E. tereticornis* growing in the vicinity of EucFACE.

*This corresponds to a 95% loss of hydraulic conductance, representing the maximum recoverable water stress in a number of species (Brodribb & Cochard, 2009; Sabot et al., 2020).

*Parameter values shared across rings appear on the right-hand-side.

**In model simulations that do not consider leaf nitrogen allocation (e.g., our default model configuration), these parameters are kept at their $t_0$ value through time.

The 36 model configurations were evaluated against observations of daily transpiration ($E$; see Section 2.4) from the aCO₂ rings only, hence the model behaviour is emergent at the eCO₂ rings. Specifically, the configurations were ranked depending on: (i) their relative variability; (ii) their similarity with the observations; and (iii) their accuracy. Relative variability was determined by the ratio of modelled to observed sample standard deviation (cf. the approach of Best et al., 2015). Similarity was measured by a skill score of overlap between modelled and observed binned distributions (Perkins et al., 2007). Model accuracy was gauged via the Absolute Scaled Error (Hyndman & Koehler, 2006), which establishes whether model predictions are more skilful than one-step
forecasts of the observations. Mathematical formulations are given in Methods S7.

External contributions to change in modelled \( V_{cmax}^{25} \) and \( J_{max}^{25} \) from the LAI, PPFD, \( T_a \), \( D_a \), \( C_a \), and the modelled soil moisture (soil water potential, \( \Psi_s \)) were analyzed for the best-ranked model configuration. To do so, we used Dominance Analysis (Azen & Budescu, 2003) which generalizes \( R^2 \) to all possible combinations of predictors before decomposing it for each predictor (i.e., environmental driver). Here, the method serves as an alternative to linear regression, to more meaningfully identify the relative contributions of each environmental driver to the variance in modelled \( V_{cmax}^{25} \) and \( J_{max}^{25} \).

2.3.3 | Foliage response to drought

Finally, we carried out an experiment to infer a link between foliage response to drought (i.e., leaf abscission and/or a lack of new leaf production, without distinction between the two processes) and plant hydraulic status. We asked: how different would plant hydraulic status have been, had LAI not declined during drought? For each ring, we computed the average LAI phenology from 2013 to 2020, and we used it to force the best-ranked model configuration. The resulting predicted PLC (from the average phenology) was compared to the PLC predicted using the observed LAI, and the difference between the two simulations shed light on the hydraulic damage avoided through foliage adjustment.

We also estimated the carbon allocated to leaf growth was then estimated from NPP using site-specific carbon allocation fractions (Table S3).

By probing the carbon benefits of increasing hydraulic risk by increasing LAI and, conversely, the carbon risks of drought-induced LAI drops, we gain a fuller understanding of how LAI dynamics relate to plant carbon and water status.

2.4 | Observational data sets

Observed \( V_{cmax}^{25} \), \( J_{max}^{25} \), and \( R_d^{25} \) were used to parameterize the model (Table 1; see Methods S6), and to evaluate the temporal changes in photosynthetic capacities predicted by the optimal leaf nitrogen allocation scheme. The observed estimates were obtained from analyzing 323 repeated light- and temperature-controlled photosynthesis-CO2 response (\( A-C_i \)) curves, using the "plantecophys" R package (Duursma 2015). \( A-C_i \) curves were measured on mature (>90 days old) and newly flushed (<90 days old) leaves of the upper canopy of three to four dominant or codominant trees in each ring, from February 2013 to February 2020 (16 measurement campaigns in spring, summer, and autumn, extending previous compilations by Ellsworth et al., 2017 and Wujeska-Klause et al., 2019).

Observations of foliar nitrogen concentration on either a mass or an area basis (i.e., mg g\(^{-1}\) or mg m\(^{-2}\)) are included for a qualitative comparison with the modelled \( N_p \). Nitrogen concentrations were measured for 236 upper canopy leaf samples that were collected from the three to four (co-) dominant trees of each ring, from February 2013 to February 2018 (11 measurement campaigns; Gherlenda et al., 2016a; Wujeska-Klause et al., 2019).

Morning (9:30–11:30 h) and afternoon (13:00–15:00 h) \( \Psi_l \) measured from October 2012 to January 2014 (Gimeno et al., 2016) were used to parameterize \( k_r \) (Table 1; Methods S6) and to evaluate
the model simulations of Ψi. The parameterization of ni was established from the three to four (co-)dominant trees in each ring, whereas the model evaluation considered 493 measurements of Ψi made on 37 trees (4–7 trees per ring).

We used tree-level E estimates for parameterization of k_{max} (Table 1; Methods S6). These daily E estimates were derived from tree sap velocities measured between January 2013 and September 2014, using the heat pulse compensation technique in the three to four (co-)dominant trees of each ring (Gimeno et al., 2018). The measurements were converted to tree-level E using ring-specific sapwood area per unit ground area ratios, as estimated from trees neighbouring the experimental rings. For comparison with our model simulations at the ring-level, we multiplied the tree-level E estimates by their respective ring’s LAI.

Finally, root-zone soil moisture was estimated from two sets of observations described in Gimeno et al. (2018). First, daily measurements were retrieved from January 2013 to December 2019, using theta probes to a depth of 30 cm at each ring. Second, soil moisture was monitored at 25 cm intervals from 25 to 150 cm depth, and at 50 cm intervals from 150 to 450 cm depth, every 10–21 days (less frequently in 2017) between January 2013 and July 2019, using neutron probes at two locations per ring. We weighted the observed soil moisture by root fraction at depth (see Methods S2 and S6) to obtain an estimate of root water access.

3 | RESULTS

3.1 | Timescales of adjustments to function

To understand whether competing optimization schemes, perturbed by hydraulic legacies, varied predictably and coherently, we examined differences between the default model configuration (optimization of canopy gas exchange only) and the 35 other configurations that combine functional adjustments. We focussed on analyzing model configuration differences on average water use efficiency (WUE; the ratio of A_{ni} to E) because WUE integrates information on both the canopy carbon and water fluxes. Additionally, we have a theoretical understanding of how WUE ought to behave across axes of variation in LAI, soil moisture, and atmospheric [CO_{2}], and this helps assess the predictability and coherence of our combined competing optimality criteria.

Figure 2 qualitatively illustrates the relative effects of differences in WUE between the default model configuration and all model other configurations, hereafter referred to as ∆WUE_{config}. These effects are shown for each of the aCO_{2} (panel a) and eCO_{2} (panel b) rings, and the rings appear in order of increasing soil moisture × LAI from left to right within each panel. Negative ∆WUE_{config} were simulated at each ring except R5, which indicated that the combinations of competing optimization schemes and hydraulic legacies mostly led to a reduction in WUE compared to the default model. However, overall, the effects were smaller for the eCO_{2} than the aCO_{2} rings (as shown by smaller relative horizontal span for the eCO_{2} rings), implying that the eCO_{2} stimulation of A_{ni} (and hence WUE) buffered WUE loss relative to the default model. Departure from the default model configuration also decreased as the soil moisture and LAI increased, with R5 (eCO_{2} ring with relatively high soil moisture × LAI) the only ring where combined functional adjustments led to increased WUE (∆WUE_{config} > 0).

Within-ring WUE changes (δWUE_{config} relative to the default WUE: see annotations in Figure 2) were small, ranging between −4.8% and 1.8% across rings. At most rings, the greatest change was associated with combinations of short timescale adjustments: the combined 14-day H_{lag} and 7-day N_{opt}. Thus, the higher the frequency of adjustments, the less water-use efficient the canopy gas exchange scheme. This result is explained by the fact that impairments from short episodes of water stress have most bearing on high frequency (short timescale) adjustments to function. For example, a 3-day reduction of C_{i} caused by soil moisture stress ought to reflect in a weekly adjustment but not in a monthly adjustment of N_{p}. The contrary was true for R5 (∆WUE_{config} > 0) where the largest gain of WUE corresponded to combinations of low frequency (long timescale) adjustments (H_{lag} = 180 days, N_{opt} = 21 days).

In all cases, N_{opt} drove change in WUE, as H_{lag}’s contributions to ∆WUE_{config} were small. Indeed, the effects from H_{lag} varied alone (e.g., H_{lag} = 14 days, at the bottom of each plot) added to the effects from N_{opt} varied alone (e.g., N_{opt} = 7 days, at the bottom of each plot) did not diverge from the effects of H_{lag} and N_{opt} varied jointly (e.g., H_{lag} = 14 days, N_{opt} = 7 days, at the top of each plot), as seen from the lack of difference between the vertical line markers (added singular effects) and the horizontal bars (interactive joint effects).

The 36 model configurations’ ability to simulate daily E at the aCO_{2} rings is reported in Figure 3. For the accuracy metric (see Methods S7), the lower the values, the better the performance; values <1 indicate that model predictions are more skilful than one-step-ahead forecasts using the preceding observations. N_{opt} alone and N_{opt} combined with the longest timescale of H_{lag} resulted in the highest accuracy (first 10 absolute ranks <0.784). For the similarity and variability metrics (see Methods S7), the closer to 1 the better. The performance was typically higher for the higher frequency H_{lag} and N_{opt} (>0.795 and >0.877, respectively), which operated more functional adjustments to short episodes of water stress than lower frequency H_{lag} and N_{opt}.

Quantile ranks were used to summarize skill across statistical metrics, considering a configuration’s performance relative to that of all other configurations within each metric. Perhaps counter intuitively, the best quantile-ranked configurations could occupy low or average positions for the accuracy metric (see the annotations in Figure 3) because there were only small differences between the 1st and 30th ranked configurations (spanning less than half the y-axis). Overall, H_{lag} alone only marginally improved on the default configuration, and we were unable to establish clear timescales of influence. By contrast, skill distinctly increased with the frequency of N_{opt} optimization (higher skill for N_{opt} = 7–14 days and lower skill for N_{opt} = 28–35 days). The best model configuration combined H_{lag} = 30 days with N_{opt} = 7 days.
3.2 | Leaf- and stand-level dynamics

Figure 4 shows that \( V_{\text{cmax}} \) and \( J_{\text{max}} \) were well captured (averages within ±5% of the observations) by the best model configuration (\( H_{\text{leg}} = 30 \) days, \( N_{\text{opt}} = 7 \) days), for both the aCO2 (a and c) and eCO2 (b and d) rings. Assessing the ability of the model to reproduce the observed dynamics was complicated by tree-to-tree variability in the measured data, particularly \( V_{\text{cmax}} \) (e.g., ranging between 33.3 and \( 116.6 \mu \text{mol C m}^{-2} \text{s}^{-1} \) at R2 on 30 January 2014). Nevertheless, the model predicted the overall observed seasonal shift from low \( J_{\text{max}} \) in spring/summer (low LAI; Figure S1) to high \( J_{\text{max}} \) in autumn (LAI peaks in autumn/winter). The model also predicted the increase in the ratio of \( J_{\text{max}} : V_{\text{cmax}} \) between summer and autumn. However, it exaggerated the decline in \( J_{\text{max}} : V_{\text{cmax}} \) between autumn and spring (i.e., between the highest and the lowest yearly LAI) thus simulating lower \( J_{\text{max}} : V_{\text{cmax}} \) in spring than summer, which contradicted the observations.

In the springs of 2013 and 2016, LAI was very low (c. 0.6 m\(^2\) m\(^{-2}\) and below 0.5 m\(^2\) m\(^{-2}\) in some of the rings; Figure S1) and the simulated \( V_{\text{cmax}} \) was overestimated due to underestimated \( J_{\text{max}} : V_{\text{cmax}} \). The simulated \( V_{\text{cmax}} \) appeared high in the springs and summers of 2017/2018 and 2019/2020 too; but there is a paucity of observations from mid-2017 onwards. Statistical analysis revealed LAI to be the main driver of seasonality in the modelled \( V_{\text{cmax}} \) and \( J_{\text{max}} \). At most rings (R2 and R3, aCO2; R1 and R4, eCO2), >74% of the modelled variance in \( V_{\text{cmax}} \) and \( J_{\text{max}} \) could be explained by variation in LAI (contributing 54%) and the environmental forcing (\( D_0 \) 9%, PPFD 6%, other drivers <5% each). At R6

![Figure 2](https://example.com/figure2)

**Figure 2** The effect sizes associated with different combinations of simulated adjustments to vegetation function, at the aCO2 (a) and eCO2 (b) rings. The horizontal bars reflect magnitude differences on average water use efficiency (WUE; the ratio of net carbon assimilation to transpiration) between the default model configuration and the model configurations that consider additional functional adjustments. There are 35 such configurations, accounting for: hydraulic legacies (\( H_{\text{leg}} \); singular effects), leaf nitrogen allocation (\( N_{\text{opt}} \); singular effects), and combinations of \( H_{\text{leg}} \) and \( N_{\text{opt}} \) (joint effects) over their respective timescales of application (days; y-axis on the left side of the figure). The vertical black lines superposed on the joint effects show the corresponding singular \( H_{\text{leg}} \) and \( N_{\text{opt}} \) effects added together (i.e., vertical lines on both sides of the horizontal bar for the joint \( H_{\text{leg}} = 14 \) days, \( N_{\text{opt}} = 7 \) days correspond to the singular \( H_{\text{leg}} = 14 \) days added to the singular \( N_{\text{opt}} = 7 \) days). There is no x-axis or scale on the figure: the horizontal span dedicated to each ring corresponds to its magnitude effect size relative to that of all other rings. The sign of the effects at each ring is indicated by an arrow (↘ for a decrease vs. ↗ for an increase in model configuration WUE), with asterisks marking instances where the sign of the effect opposes all other effect signs within a ring (e.g., \( H_{\text{leg}} = 180 \) days at R6). At each ring, the largest percentage change in model configuration WUE is indicated over its corresponding functional adjustments. Finally, within each panel, the rings are ordered in increasing order of soil moisture × leaf area index from left to right. [Color figure can be viewed at wileyonlinelibrary.com]
(aCO₂) and R5 (eCO₂), both LAI and soil moisture were typically higher than in the other rings and the environmental drivers explained less of the variance in the modelled $V_{\text{cmax}}^{25}\text{J}_{\text{max}}^{25}$ (c. 61%). The contribution of LAI was not as substantial (31%), whilst that of soil moisture ($\Psi_s$) increased to 17%.

At the eCO₂ rings, the down-regulation of $V_{\text{cmax}}^{25}$ averaged to 9.6% in the observations versus 11.2% in the simulations (Table 2). In the model, this down-regulation of $V_{\text{cmax}}^{25}$ partly originated from the initialization of $N_P$ using ring-specific $V_{\text{cmax}25\text{t}}$ and $J_{\text{max}25\text{t}}$ (see Table 1 and Methods S6). When $N_P$ was initialized using the average $V_{\text{cmax}25\text{t}}$ and $J_{\text{max}25\text{t}}$ from across the aCO₂ rings (Figures S3 and S4), our leaf nitrogen allocation scheme only predicted a 5.5% reduction in $V_{\text{cmax}}^{25}$ at the eCO₂ rings. This underestimation of the eCO₂ down-regulation of $V_{\text{cmax}25}$ was explained by the model's overestimation of $V_{\text{cmax}25}$ at low LAI, i.e., at R1 and R4 (see Table 2). At R5, LAI was comparatively high, and it did not drive $V_{\text{cmax}25}$ as strongly; consequently, the model successfully down-regulated $V_{\text{cmax}25}$.

Notwithstanding initialization, the dynamics of $N_P$ (Figures S2 and S4) aligned with those of the observed leaf nitrogen concentration on an area basis, albeit displaying one order of magnitude less variability. Finally, similarly to the observed leaf nitrogen concentration on a mass basis, $N_P$ was lower at the eCO₂ rings than at the aCO₂ rings.

Figure 5 shows comparisons between the GPP, $E$, and the canopy PLC simulated by the default and the best model configurations. The functional adjustments of the best model configuration had little effect on the fluxes over time, with c. −4.8% to +1.8% GPP across rings (akin to the within-ring WUE changes from Figure 2) and $E$ virtually unchanged (−0.5% to +1.8%; $r^2 = 0.7$ in each case). Magnitude differences in simulated GPP between the default and the best model configurations were the
FIGURE 4 Timeseries of the $V_{\text{cmax}_{25}}$ (a and b) and $J_{\text{max}_{25}}$ (c and d) predicted by the best model configuration ($H_{\text{leg}} = 30$ days, $N_{\text{opt}} = 7$ days) and compared with field observations at the aCO2 (a and c) and eCO2 (b and d) rings. Orange lines show the average simulations across rings (e.g., average of R2, R3, and R6 for the aCO2 rings) and yellow shadings the range of simulations. Box and whisker plots summarize the observed distributions of $V_{\text{cmax}_{25}}$ and $J_{\text{max}_{25}}$ (line, median; box, interquartile range; whiskers, quartiles ± 1.5 times the interquartile ranges) during measurement campaigns, with each observation displayed via a dot. Fixed parameter values appear in each panel, these refer to the ring-specific parameterizations used by the default model configuration, as well as in the initialization of the leaf nitrogen allocation scheme.

TABLE 2 A comparison of observed and simulated maximum carboxylation rates at 25°C ($V_{\text{cmax}_{25}}$; μmol C m$^{-2}$ s$^{-1}$)

<table>
<thead>
<tr>
<th></th>
<th>aCO2</th>
<th>eCO2</th>
</tr>
</thead>
<tbody>
<tr>
<td></td>
<td>R2</td>
<td>R3</td>
</tr>
<tr>
<td>Observations*</td>
<td>92.3</td>
<td>92.1</td>
</tr>
<tr>
<td>Simulations</td>
<td></td>
<td></td>
</tr>
<tr>
<td>Parameterization**</td>
<td>92.7</td>
<td>92.1</td>
</tr>
<tr>
<td>Model predictions***</td>
<td>96.1</td>
<td>94.7</td>
</tr>
<tr>
<td>Parameterization****</td>
<td>93.8</td>
<td></td>
</tr>
<tr>
<td>Model predictions****</td>
<td>95.7</td>
<td>98.2</td>
</tr>
</tbody>
</table>

Note: The aCO2 (R2, R3, and R6) and eCO2 (R4, R1, and R5) rings appear ordered from smallest to highest average soil moisture × leaf area index, as in Figure 2. The percentage down-regulation of $V_{\text{cmax}_{25}}$ at eCO2 is shown in italic below the average aCO2 | eCO2 pairing. Two sets of simulations conducted with the best model configuration ($H_{\text{leg}} = 30$ days, $N_{\text{opt}} = 7$ days) appear in the table: (i) simulations initialised by ring-specific parameterizations of $V_{\text{cmax}_{250}}$ and $J_{\text{max}_{250}}$; (ii) additional simulations initialised by the average $V_{\text{cmax}_{250}}$ and $J_{\text{max}_{250}}$ from across the aCO2 rings applied to all rings. For context, the parameterizations of $V_{\text{cmax}_{250}}$ appear above the modelled values of $V_{\text{cmax}_{25}}$.

*Observational averages are different from the parameterizations because the parameter values derive from campaign averages (equal weights for each campaign; see MethodsS6) whereas these averages apply equal weights to each date of collection.

**These parameter values are the same as the $V_{\text{cmax}_{250}}$ presented in Table 1.

***Simulation averages only include predictions at dates for which observations are available.

****The average $V_{\text{cmax}_{250}}$ from across the aCO2 rings (i.e., R2, R3, and R6) is used for all rings.

largest in spring/summer and fluctuated annually (e.g., $-5$ g C m$^{-2}$ y$^{-1}$ $[-0.3\%]$ vs. $-86$ g C m$^{-2}$ y$^{-1}$ $[-11.7\%]$ at R2 in 2014 [high LAI year] versus 2018 [low LAI year]). Even when GPP was markedly reduced by the best model configuration, the canopy gas exchange scheme kept $E$ roughly level with the default model configuration, which limited further drops in GPP but reduced WUE.

Departure by the best model configuration from the default was more pronounced at the aCO2 rings ($-72$ g C m$^{-2}$ y$^{-1}$ GPP; $-2.3$
mm y⁻¹ E) than at the eCO₂ rings (−42 gC m⁻² y⁻¹ GPP; +0.0 mm y⁻¹ E). With this disproportionate aCO₂/eCO₂ response between the model configurations, the baseline relevant to estimating the simulated CO₂ fertilization effect on GPP changed and, as a result, the best model configuration simulated a greater CO₂-driven increase in GPP than the default model configuration (+13.1% GPP at the eCO₂ rings relative to the aCO₂ rings, compared with +9.6% GPP). Notably, the simulated eCO₂ increase in WUE was not proportional to the increase in GPP in either model configuration (+9.7% WUE and +6.8% WUE in the best and default configurations, respectively), because E was higher at the eCO₂ than at the aCO₂ rings (see Discussion).

Our model captured Ψᵢ well (insets of Figure 5e,f). Marked canopy PLC (Figure 5e,f) was simulated in the second half of 2014, in mid-2016, and in late 2017 (Ψ₉₉ > 45% vs. < 10% in 2015, 2018, and 2019), in association with a lack of sustained hydraulic legacies. That is, hydraulic legacies did not last more than 2 months and reaching a maximum ΔPLC < 15% between the canopy PLC of the best (legacies) and default (no legacies) model configurations (Figure 5e,f). This lack of sustained hydraulic legacies could be assumed to have arisen from an overestimation of soil moisture, especially after 2017 (Figure S5). Nevertheless, hydraulic legacies remained small when soil moisture was prescribed from observations (Figure S6). Possible explanations for these limited hydraulic legacies and the ensuing lack of effect on the surface fluxes are: (i) that changes from legacies compensated each other (Figure S7) because diel and day-to-day fluxes were highly dynamic; and/or (ii) that legacies were realized though extra-xylary adjustments that allowed the trees to maintain function, for instance, adjustments to LAI.
3.3 | Foliage response to drought

Figure 6 further explores the role of LAI in response to water stress, by showing how observed deviations in LAI from the average phenology impact the simulated canopy PLC, considering associated carbon benefits and risks. Higher LAI than the average phenology generally resulted in a net PLC increase of 1.5% on top of a background PLC of 4.2% at the aCO2 rings. Higher LAI also led to a 2.9% net PLC increase on top of a background PLC of 3.6% at the eCO2 rings. In mid-2014, after a multi-year peak in LAI at all rings, up to 45% and 78% net increases in PLC were simulated from background PLCs < 5% at the aCO2 and eCO2 rings, respectively. This particularly high PLC may explain the ensuing low observed LAI in 2015.

Increased PLC from the model forced with LAI higher than the average phenology was accompanied by increasing cumulative carbon storage until either 2016 (67 gC m⁻²; eCO2 rings) or 2017 (63 gC m⁻²; aCO2 rings). After those years, the LAI declined greatly, reducing carbon uptake and storage. Still, total accumulated leaf carbon stocks were higher in 2019 than they would have been if the canopy had remained as per the average phenology (ΔyAfcum > 0 in the insets of Figure 6c,d), highlighting the beneficial nature of the observed LAI dynamics. Additionally, forcing the model with LAI lower than the average phenology deducted c. 1.8% from a background PLC of 5.8% at the aCO2 rings, and c. 2.5% from a background PLC of 5.3% at the eCO2 rings. In mid-2018, as much as 73.3% and 85.4% net reduction in PLC were simulated from respective background PLCs of 81.0% and 89.8% at the aCO2 and eCO2 rings.

On the whole, both net PLC increases from higher LAI and net PLC decreases from lower LAI were more important at the eCO2 than at the aCO2 rings (Figure S8), though the relationship was asymmetrical and favoured the net PLC decreases. This result implies higher eCO2 ratios of ΔPLC:ΔLAI that were further enhanced by drought and/or low LAI.

4 | DISCUSSION

4.1 | Representing legacies from climate and weather extremes

Our model-based approach suggests that the E. tereticornis trees at EucFACE were resilient to recurring drought (i.e., yearly soil moisture dry-down, multiyear drought after mid-2017) and heatwave events.
(i.e., temperatures >40°C and vapour pressure deficit >6 kPa every spring/summer) between 2013 and 2020. For instance, non-negligible canopy PLC (c. 9.5%–28.5%) was simulated during the spring/summer dry-down of 2013, yet both LAI (Figure S1) and the transpiration flux (Figure S5) seemingly recovered to “peak” levels by autumn 2014. Further marked canopy PLC was simulated in mid-2014, mid-2016, and late 2017, although the exact magnitude PLC emerging from our model relates to its ability to represent the surface extraction of water by roots (cf. Figures 5 and S6).

In our simulations, high canopy PLC subsided rapidly, as hydraulic legacies and their impacts on the transpiration were small. This outcome partially stemmed from our parameterization of the ratio of xylem recovery from embolism \( r_E \) (Methods S6) which, at c. 0.9 (Table 1), implies a high ability to recover lost hydraulic conductance. A recent study of hydraulic legacies in two Eucalyptus hybrids (Saunders & Drew, 2021) pointed to a possible trade-off in hydraulic strategy between vulnerability to cavitation and recovery from embolism, whereby the hybrid characterized by higher hydraulic vulnerability (faster stomatal closure) recovered conductance faster than its counterpart that better resisted the formation of embolism. This mechanism could be envisioned for the E. tereticornis examined here, as their \( P_{50} \) of −4.07 MPa confers less resistance to cavitation than the \( P_{50} \) of other trees in the vicinity of EucFACE (e.g., E. fibrosa: −5.82 MPa, E. moluccana: −5.74 MPa, Melaleuca decora: −5.68 MPa; Peters et al., 2021).

A drawback of our approach to modelling legacy is that it is sensitive to the parameterization of \( r_E \). However, xylem recovery capacity itself may be sensitive to drought severity (Gauthie et al., 2022), and so may the value of \( r_E \). Therefore, even though we derived \( r_E \) from site measurements over a period that comprised notable dry spells (October 2012–November 2013), we cannot exclude the possibility that \( r_E \) may be different during a longer, more severe drought, for example, if it were derived using data from the 2017–2020 drought period. More generally, reliance on species-specific parameterizations of \( r_E \) would increase model complexity without always impacting hydraulic capacity and/or canopy fluxes. Meanwhile, there is mounting evidence of interconnected drought- and heat-induced legacies on nonstructural carbohydrates (NSCs) and hydraulic function (Duan et al., 2022; Poyatos et al., 2013; Ruehr et al., 2019). The key to LSMs accounting for differentiated legacies from climate extremes (i.e., depending on species and/or location) might lie in a mechanistic link between NSCs and plant hydraulics. Representing NSCs (see e.g., Jones et al., 2020) alongside photosynthetic, stomatal, and hydraulic processes in LSMs would provide the opportunity to elucidate such a link, by way of observation-driven model experiments.

### 4.2 Leaf nitrogen allocation and photosynthetic capacities

Previous tests of optimal leaf nitrogen allocation schemes in LSMs (Caldararu et al., 2020; Haverd et al., 2018) had not considered water-limited, low LAI ecosystems such as EucFACE (long-term LAI of c. 0.8 m² m⁻² across rings), where GPP is predominantly limited by the ability to capture light (\( J_{\text{max}25} \)) rather than to fix CO₂ (\( V_{\text{max}25} \)). Ecosystems in dry environments play a noteworthy role as drivers of interannual variability in the global carbon cycle (Ahlstrom et al., 2015; Poulter et al., 2014), and so, the dynamics of their limiting rates to photosynthesis warrant further attention.

Our optimal leaf nitrogen allocation scheme was generally able to resolve the dynamics of \( V_{\text{max}25} \) and \( J_{\text{max}25} \) (Figure 4, but it overestimated \( V_{\text{max}25} \) when/where LAI was very low (c. 0.6 m² m⁻²; see Figure S1). The model also failed to predict the eCO₂ down-regulation of \( V_{\text{max}25} \) where the LAI was very low (i.e., R1 and R4 vs. R5 where it was successful; see Table 2). Either the model's scaling from the leaf- to the canopy-level is inadequate when LAI is very low, or the leaf nitrogen allocation routine wrongly works to correct for the LAI reduction during drought, or both. In alternative work predicting seasonal variations in \( V_{\text{max}25} \) from an optimality principle, Jiang et al. (2020a) argued the need to account for site-specific coefficients of nitrogen extinction within the canopy (\( \kappa_N \)) correlated with LAI. With help from existing observations of leaf nitrogen, \( V_{\text{max}25} \) and \( J_{\text{max}25} \) depending on canopy height (e.g., Crous et al., 2021) for EucFACE), future studies may evaluate whether varying \( \kappa_N \) reduced low-LAI simulation biases.

Errors in the simulation of \( V_{\text{max}25} \) at low LAI depended on whether the amount of leaf nitrogen available for photosynthesis, \( N_p \), was initialised at the ring-level (see Table 2). However, the simulated dynamics of \( N_p \) aligned with observed adjustments in leaf nitrogen regardless (Figures S2 and S4). \( N_p \) declined under elevated [CO₂], due to a simulated reduction in the soluble protein nitrogen invested in Rubisco concomitant to an increase in other soluble protein investments. This shift in the nitrogen invested into photosynthetic compounds is echoed by observational evidence for the upper canopy of E. globulus grown at the Hawkesbury Forest Experiment near EucFACE (Sharwood et al., 2017). The simulated decline in \( N_p \) at the eCO₂ rings was also in qualitative agreement with the observations of leaf nitrogen concentration on a mass basis (cf. Crous et al., 2019; Wujeska-Klause et al., 2019), but not with the observations of leaf nitrogen on an area basis (i.e., accounting for variation in leaf mass per area) that were more often than not higher or level with the aCO₂ rings. This mismatch between the behaviour of the leaf nitrogen concentration on a mass versus an area basis, along with the sensitivity of our modelled \( V_{\text{max}25} \) to the initialization of \( N_p \), suggests drought/ LAI-driven reallocations of plant nitrogen at the eCO₂ rings consistent with the findings of Crous et al. (2019). Exploring these reallocation patterns is beyond the scope of the current study, but could be done exploiting theoretical frameworks that link nitrogen and water balance to LAI (e.g., McMurtrie et al., 2008).

### 4.3 Why did combined model adjustments to function reduce water-use efficiency?

Many studies that consider optimality principles allow plants to reach optimum points without accounting for potential ‘suboptimality’ in
behaviour (Caldararu et al., 2020). A novel aspect of our study is that we formalized our optimality criteria such that they interact dynamically whilst not serving the same purpose, and we also limited optimal behaviour via a representation of hydraulic legacies. Together, the interacting optimality criteria and hydraulic legacies may mimic diversity in the potential causes of plant plasticity as well as phenotypic limitations to plasticity (Grams & Andersen, 2007; Grassi & Bagnaresi, 2001; Merliå & Hendry, 2014; Stotz et al., 2021).

Our two optimality criteria interacted coherently, including when paired with hydraulic legacies. Their combinations showed predictability with respect to environmental factors (soil moisture, LAI, [CO2]), and behavioural coherence over their timescales of application (Figures 2 and 3). Their propagation to the stand-scale was small (i.e., <5% average decline in GPP), but of similar magnitude to the findings of other studies that considered seasonal changes in photosynthetic capacities (Bonan et al., 2011; Medvigy et al., 2013). Here, it is also worth noting that our comparison point (i.e., the default model configuration) is a canopy gas exchange optimization approach that has been shown to significantly outperform standard LSM implementations (Sabot et al., 2020).

Contrary to leaf-level photosynthesis where Rubisco limitations on photosynthesis prevail throughout the day, our stand-scale GPP simulations were chiefly limited by RuBP regeneration, owing to relatively low ratios of \( J_{\text{max25}} : V_{\text{cmax25}} \) partly caused by the low LAI at EucFACE (cf. predictions by the forest canopy model MAESPA; Yang et al., 2020). Due to this RuBP regeneration limitation, \( J_{\text{max25}} \) is a much more relevant predictor of GPP than \( V_{\text{cmax25}} \) at EucFACE, particularly when LAI drops during drought (at the end of 2013, from 2018 onwards; see Figures 4 and 5). Decreased GPP from the down-regulation of \( J_{\text{max25}} \) translated into quasi-equivalent changes in WUE, because our optimization model sought to maintain \( E \) (preventing excessive reductions of \( C_v \)) to prevent the further decline of GPP during low LAI/drought periods. Unlike the transpiration simulated here, a ‘standard’ LSM (e.g., CABLEv2.0) would have predicted \( E \) to decline with the simulated reduction in GPP, by keeping the WUE quasi-constant.

One potentially influential physiological association that we did not examine is the link between \( k_{\text{max}} \) and \( V_{\text{cmax25}} \). This association was explored by Sabot et al. (2020) over decadal timescales and their results broadly supported a link between \( k_{\text{max}} \), average \( V_{\text{cmax25}} \), and long-term climate. In theory, \( k_{\text{max}} \) should vary with the length of the hydraulic path between the roots and the leaves, so it ought to adjust (sub-seasonally as photosynthetic investments shift to different areas of the canopy (e.g., from the top of the sunlit canopy to lower shaded leaves; Peltoniemi et al., 2012). Still, whether \( k_{\text{max}} \) adjusts in coordination with \( V_{\text{cmax25}} \) and/or \( J_{\text{max25}} \) over timescales as short days or weeks, and whether that is to maximize GPP or to preserve WUE remain open questions.

### 4.4 Implications at elevated [CO2]

At eCO2, our best model configuration (i.e., high-frequency combinations of optimality criteria) predicted a 13.1% increase in GPP (+3.5 percentage points compared to our default model), which compared well with a model-data fusion estimate of a 12% increase in GPP over the 2013–2016 period (Jiang et al., 2020b).

Both measurements of leaf-level stomatal conductance (Gimeno et al., 2016) and our model predictions of canopy conductance agree on a eCO2-induced decline of the order of 20% under favourable conditions (mid-morning, no water stress). At the same time, both observations of tree-level transpiration (Gimeno et al., 2018) and our modelled stand-scale transpiration were not reduced by eCO2. Instead, they were slightly higher under eCO2. This apparent contradiction is explained by a higher average soil-to-canopy maximum hydraulic conductance under eCO2 \( (k_{\text{max}} = 1.44 \text{ mmol m}^{-2} \text{s}^{-1} \text{MPa}^{-1}) \) than aCO2 \( (k_{\text{max}} = 1.16 \text{ mmol m}^{-2} \text{s}^{-1} \text{MPa}^{-1}) \), accompanied by higher \( \psi \ell \) (c. 0.15–0.22 MPa higher), which may indicate hydraulic acclimation and/or reflect the heterogeneity of the site, for example, differences in soil textures and water retention (Table 1).

Observational evidence of hydraulic acclimation to eCO2 is equivocal (De Kauwe et al., 2021; Domic et al., 2017). Nevertheless, physiological adjustments of the same nature as those captured here (i.e., higher \( k_{\text{max}} \), higher \( E \), higher \( \psi \ell \)) have also been observed for Eucalyptus grown in a glasshouse and exposed to dry-downs and eCO2 (Jiang et al., 2021). Moreover, both higher \( k_{\text{max}} \) and the higher eCO2 ratios of \( \Delta \text{PLC}:\Delta \text{LAI} \) simulated in our “foliage response” experiment are consistent with the eco-hydrological equilibrium hypothesis (Sperry et al., 2019; Yang et al., 2018).

### 4.5 Foliage response to drought

We wanted to understand the mechanistic causes for the adjustments in LAI that are observed with changes in rainfall at EucFACE (Duursma et al., 2016). To do so, we linked LAI, carbon storage, and hydraulic status in a model, which has seldom been attempted before (but see Trugman et al., 2018; Xu et al., 2016). This approach contrasts predominantly statistical efforts to identify legacy effects at the landscape- and ecosystem scales (see Kannenberg et al., 2020) for a review), similar to most work on ecosystem resilience to repeated droughts and heatwaves (e.g., Mitchell et al., 2014; Neumann et al., 2017).

2015 was the only year for which we could potentially link low LAI to a compromised hydraulic system (i.e., marked simulated PLC in mid-2014). Coincidentally, an outbreak of senescence-feeding psyllids (Cardiaspina fiscella) caused major defoliation at EucFACE in September 2014 (Gherlenda et al., 2016b). More generally, though, we found that foliage responses to drought (i.e., leaf abscession and/or a lack of leaf growth) prevented the loss of hydraulic function (Figure 6). Drought-induced drops in LAI are often perceived as a pervasive consequence of water stress that negatively impacts carbon stocks (e.g., Anderegg et al., 2015). But at EucFACE, under water-limited conditions, an alternative interpretation is that this strategy could be akin to “ecological memory” benefitting the trees. Foliage growth was sufficient in the relatively wetter years to buffer carbon losses associated with lower LAI during the drier years (2013–2016/2017 vs. 2017–2020; see Figure 6).

These findings concur with Eucalyptus canopy dynamics observed in the field (Battaglia et al., 1998; Pook, 1985; Whitehead & Beadle, 2004), but
explicit linkage of observed hydraulic status, LAI, and carbon storage/ 
biomass remain rare (but see Atwell et al., 2007; Nadal-Sala et al., 2021; 
Poyatos et al., 2013). Future work that examines this link across species 
and ecosystems will be critical.

5 | CONCLUSION

In this study, we demonstrated that competing nitrogen-based and 
hydraulics-based optimization schemes, impacted by a representation 
of hydraulic damage, affected model simulations in a predictive 
manner along axes of variation in soil moisture, LAI, and atmospheric 
[CO₂], via interactive physiological adjustments that could also be 
interpreted theoretically. This result should be generalizable beyond 
EucFACE, opening avenues to reconcile alternative optimality 
principles and to help inform predictions of tree species’ resilience 
to climate change.

We also showed that the Eucalyptus trees growing at EucFACE 
could suffer significant canopy PLC, but this did not lead to sustained 
hydraulic damage in our model simulations. Instead, legacy effects 
were accounted for through LAI dynamics that mostly prevented 
hydraulic damage. Leaves were grown during wet years to replenish 
carbon stores, whereas foliage declined in anticipation of water 
stress. Overall, capturing these dynamics during drought appears to 
be a key research frontier: models that ignore or do not accurately 
represent LAI responses to water availability risk predicting pre-
mature canopy dieback and/or plant mortality.

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CODE AND DATA AVAILABILITY STATEMENT

All model, analysis code, and data files are freely available from 
https://doi.org/10.5281/zenodo.6717290 (Sabot, 2022) and the code 
is also available from https://github.com/ManonSabot/Competing_ 
Optimal_Adjustments. Previously published data sets used in this study 
can be accessed at: 
http://doi.org/10.4225/35/57ec5d4a2b78e (Ellsworth et al., 2017). 
http://doi.org/10.4225/35/55b6ec313444f (Gimeno et al., 2016). 
http://doi.org/10.4225/35/5ab9bd1e2f4fb (Gimeno et al., 2018).

ORCID

Manon E. B. Sabot http://orcid.org/0000-0002-9440-4553 
Martin G. De Kauwe http://orcid.org/0000-0002-3399-9098 
Andy J. Pitman http://orcid.org/0000-0003-0604-3274 
David S. Ellsworth http://orcid.org/0000-0002-9699-2272 
Belinda E. Medlyn http://orcid.org/0000-0001-5728-9827 
Silvia Calderaro http://orcid.org/0000-0001-5839-6480 
Sönke Zaehle http://orcid.org/0000-0001-5602-7956 
Kristine Y. Cruss http://orcid.org/0000-0001-9478-7593 
Teresa E. Gimeno http://orcid.org/0000-0002-1707-9291 
Mengyuan Mu http://orcid.org/0000-0001-6517-5504 
Jinyan Yang http://orcid.org/0000-0002-4936-0627

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SUPPORTING INFORMATION

Additional supporting information can be found online in the Supporting Information section at the end of this article.