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1 Title:

2 **Long-term accumulation and transport of anthropogenic phosphorus in three river basins**

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24 Abstract:

25 Global food production depends on phosphorus (P) and in agricultural and urban
26 landscapes much P is anthropogenically cycled through trade. Here we present a long-term,
27 large-scale analysis of the dynamics of P entering and leaving soils and aquatic systems via a
28 combination of trade, fluvial transport and waste transport. We report net annual P inputs, and
29 the P mass accumulated over several decades, for three large river basins. Our analyses reveal
30 historical P accumulation for two mixed agricultural-urban landscapes (Thames Basin, UK;
31 Yangtze Basin, China), and one rural agricultural landscape (Maumee Basin, USA). We also
32 show that human modes of P transport involving trade and waste massively dominate over
33 fluvial transport in these large basins, and we illustrate linkages between fluvial P dynamics and
34 infrastructure such as wastewater treatment and dams. For Thames and Maumee Basins, recently
35 there was modest P depletion/drawdown of the P pool accumulated in prior decades, whereas
36 Yangtze Basin has consistently accumulated P since 1980. These first estimates of the magnitude
37 of long-term, large-scale P accumulation in contrasting settings illustrate the scope of
38 management challenges surrounding the storage, fate, exploitation and reactivation of legacy P
39 that is currently present in the Earth's critical zone.

40

41 Phosphorus (P) is a key requirement for food production and over the past 75 years,
42 agricultural demand has increased the rate of global P mobilization four-fold¹⁻³. Inefficiencies
43 and large losses of P occur at many points in food production, and the great majority of P
44 fertilizer originates in mines^{4,5}, raising concerns about long-term supplies of affordable fertilizer
45^{6,7}. Fluvial transport of P from agricultural land, and release of P-rich animal and human wastes
46 into the environment, have degraded lakes, rivers, reservoirs, and coastal waters with excess P,

47 causing costly damages ^{8,9}. These widespread inefficiencies in human P use have been
48 characterized as a wholesale disruption of the global P cycle ⁶ that for ages has supported
49 biological productivity through efficient recycling of P.

50 Phosphorus inputs to agriculture initially increase soil fertility and crop yields, but
51 continued P application in excess of plant uptake increases the risk of P loss from land to water
52 bodies. Following storage in soils and aquatic sediments, the associated time lags for P
53 mobilization and transport can last years to decades ¹⁰⁻¹². This relates to the notion that streams
54 have a chemical memory of the past ^{13,14} that delay recovery from water quality impairment.
55 There have been few long-term studies of the landscape-level storage, transport, and fate of P
56 accumulated in human-dominated basins (but see ^{8,12,15-17}), although there has been much
57 research on P in large basins over shorter time frames ¹⁸. Similarly, there have been few direct
58 comparisons of fluvial vs. human modes of P transport at broad scales (but see ¹⁹). Rather, much
59 research on P has involved studies of relatively short-term processes at the plot scale or within
60 individual ecosystems. This reflects the long-standing problem that changes in landscape-level P
61 storage and legacy P are very difficult to measure directly. To address these needs, we
62 synthesized diverse agronomic, urban, and river data sets, and examined the long-term dynamics
63 of P accumulation in three large river basins using a difference approach. In advance of our
64 calculations for long-term P accumulation, we also examined the dynamics of component P
65 flows involving trade, fluvial transport, and waste transport (food waste disposal, sewer
66 infrastructure) which have not been frequently juxtaposed over the long-term at large scales.

67 Our synthesis of long-term P fluxes involves: cropland-dominated Maumee River basin,
68 USA, tributary to Lake Erie, southernmost of the Laurentian Great Lakes; mixed agricultural-
69 urban Thames River basin, UK, which drains parts of the London metropolitan area *en route* to

70 the North Sea; Yangtze River basin, the largest in China, which has undergone rapid population
71 growth and economic development. To conceptualize these broad-scale P dynamics, Haygarth et
72 al.²¹ recently hypothesized that human-dominated catchments consist of an accumulation phase,
73 when P gradually builds up, and a depletion phase (Fig. S1, Supplementary Information), when P
74 inputs decline and mobilization of accumulated “legacy” P becomes an increasingly important
75 consideration. Here, we test this accumulation-depletion hypothesis, posing three questions: 1)
76 Which P fluxes drive the long-term dynamics in human-dominated river basins? 2) How do gross
77 P inputs and outputs, and net P inputs, change over the long-term? 3) How can understanding of
78 long-term accumulation inform management of P trajectories, regionally, nationally, and
79 internationally? The Maumee, Thames, and Yangtze Basins differ substantially in terms of socio-
80 economic history and physiographic features but are linked by common interests of water
81 security, food security, and resource management that transcend geopolitical hierarchies and
82 provide lessons about P.

83 Biogeochemical studies of watersheds and landscapes commonly focus on fluvial fluxes
84 but, in the Anthropocene, the P cycle has become increasingly dominated by human fluxes via
85 trade of fertilizer and food as well as management of food waste and sewage. Our analysis
86 provides new evidence that, indeed, human P fluxes massively dominate over the fluvial fluxes,
87 even for large basins. In the agricultural Maumee Basin, both annual fertilizer P import and
88 food/feed P export exceeded fluvial P export by 5- to 20-fold (Fig. 1), depending on the year. In
89 the Thames Basin, between World War II (1940) and 1980, fertilizer P import averaged >15-fold
90 higher than river P export; food/feed P export from farms >7-fold higher; food waste P to
91 landfills >4-fold higher; and P input from sewage treatment works >2-fold higher. Likewise,
92 even during the era of highest sewage P effluent and highest river P export in Thames Basin

93 (1970-1990), mean fertilizer P import, food/feed P export from farms, total sewage production,
94 and food waste P to landfills were 11, 8.0, 4.0, and 3.3 kilotons (kt) per year, respectively,
95 compared to only 1.9 kt yr⁻¹ for river P export. These results for Maumee and Thames Basins
96 suggest the changes in global fluxes of P since pre-industrial times may rival or exceed the
97 changes in the global fluxes of N and C that have been reported^{1, 21}. These major human
98 alterations to the global P cycle are compatible with previous findings for heavier elements²²,
99 whose pre-industrial cycles in the biosphere were controlled mainly by rock weathering but now
100 are being mobilized more rapidly from the crust via mining.

101 In the Yangtze River, dissolved P export increased by 10-fold between 1970 and 2010
102 but our calculations indicate a 44% decline in river total P export between 1970 and 2010
103 ($p < 0.001$, Fig. S5). This reflects a long-term decline in particulate P export that is likely linked to
104 lower suspended sediment following the construction of large dams²³, possibly combined with
105 improvements in sewage treatment. Nonetheless, like the Maumee and Thames, total P transport
106 in the Yangtze River was dwarfed by annual fertilizer P application, which increased by more
107 than 10-fold over this period of record. We suggest the dominance of human P fluxes over
108 fluvial fluxes extends to many other agricultural and urban basins of the world.

109 The highly agricultural Maumee Basin is the primary source of P to Lake Erie, where the
110 return of major algae blooms in summer 2014 resulted in the shutdown of the drinking water
111 supply to Toledo, Ohio²⁴. Prior to 1990, and as previously shown²⁵, gross P input greatly
112 exceeded gross output (Fig. 2), consistent with expectations for P accumulation (Fig. S1). Since
113 the late 1990s, gross P input and output have converged towards a common value between 15
114 and 20 kt yr⁻¹. Our analyses reveal that inter-annual variations in gross P input and output in the
115 1990s and 2000s had only a minor influence on the >200 kt pool of P that accumulated mostly

116 during the 1970s and 1980s (Fig. 3). While annual P output has exceeded input for certain years
117 (1997-1998, 2006, 2009), our calculations up to 2010 indicate there has not yet been meaningful
118 P depletion.

119 Unlike Maumee Basin, the Thames Basin includes a substantial human population
120 including parts of London. Nevertheless, akin to the Maumee, gross P input to the Thames Basin
121 greatly exceeded output until the 1990s, demonstrating a prolonged phase of P accumulation.
122 Since the late 1990s, gross annual P outputs from the Thames Basin have slightly exceeded the
123 inputs. During the 2000s, Thames River P export declined by 86 % ($p=0.001$) in association with
124 a reduced flux from sewage treatment to river, reflecting higher sewage treatment efficiency
125 motivated partly by the European Union's Urban Waste Water Directive. Over the same recent
126 period, fertilizer P import declined by 26% ($p<0.001$), while food/feed P export increased by
127 22% ($p=0.044$). Thus the Thames Basin shifted to modest depletion around 1998, following a
128 long-term decline in fertilizer P import that began around 1960 (Fig. 1 and 3).

129 In contrast to the slowing rates of P accumulation in Maumee and Thames Basins, the
130 available P data for Yangtze reveal a consistent phase of rapid P accumulation, especially since
131 1980. We were unable to determine Yangtze Basin sewage inputs ($P_{sewage,in}$) or exports of food
132 and feed ($P_{food/feed,out}$) needed in Eq. 5 (Supplementary Information), so we did not estimate gross
133 P input and output for this basin. Nevertheless, we provide estimates of net P input based on the
134 assumption of $P_{sewage,in} = P_{food/feed,out}$. Our calculations reveal that Yangtze Basin, one of Earth's
135 largest, was accumulating legacy P at a remarkable rate of 1.7 Tg yr^{-1} (1700 kt yr^{-1}) in 2010 (Fig.
136 3). On an areal basis, Yangtze Basin net annual P input of $940 \text{ kg km}^{-2} \text{ yr}^{-1}$ in 2010 approaches
137 the maximum historical rate of P accumulation in Maumee Basin ($1300 \text{ kg km}^{-2} \text{ yr}^{-1}$ in 1981) and
138 exceeds the maximum historical rate of Thames Basin ($820 \text{ kg km}^{-2} \text{ yr}^{-1}$ in 1950). This annual

139 rate of accumulation is also equivalent to about 8% of the global rate of P production from
140 phosphate rock, or 43% of the national rate of P production by China ², suggesting that Yangtze
141 Basin alone accounts for 17% of the annual P increment of 10 Tg yr⁻¹ that has been reported for
142 erodible soils globally ^{8, 12}. Like the Maumee and Thames basins, much accumulated P in the
143 Yangtze Basin occurs in arable upland soils ²⁶ and eventually could be delivered to water bodies,
144 adding to the more immediate effects of population change, dam construction, and sewage
145 treatment on dissolved or particulate P transport by rivers globally ²⁷. Research is still needed to
146 understand how interactions between land use change and climate variability affect the
147 mobilization of legacy P from soils as well as from river channels, reservoirs, floodplains,
148 wetlands, and natural lakes occurring within hydrologic networks.

149 Here we have demonstrated that large-scale assessments of landscape P storage and
150 dynamics may be achieved by difference, as previously shown in global analyses of P ^{8, 12}. This
151 approach provides a means for estimating the mass of legacy anthropogenic P that is currently
152 present in the Earth's critical zone, and may inform efforts to exploit it ⁴. Contributing challenges
153 to the direct measurement of change in P storage are that soil P is notoriously heterogeneous in
154 space and with soil depth, while historical soil sampling efforts have rarely targeted the entire
155 landscape P pool. Thus, while P flux data are often lacking during early stages of P
156 accumulation, even in intensively monitored basins such as Maumee, there are pathways for
157 long-term analysis through linkages between the P cycle and documented human activities.

158 Concerns about excess P, its mobilization, and the lack of robust P recycling pathways ^{5, 6}
159 are growing worldwide. These kinds of long-term portraits of P storage, mobilization, and
160 transfers are needed to help understand the true causes and consequences of P transport. We
161 suggest an important role for new technologies and land practices that specifically target legacy

162 P in terms of storage, fate, exploitation/recovery, and reactivation to more plant-available forms
163 ¹⁶. While our analysis has focused on a few major P-consuming nations ⁵, the need for robust P
164 recycling pathways extends to developing nations, especially those where mineral P is scarce ²⁸.
165 In regions of intense P surplus ²⁹, managed drawdown of excess soil P represents an increasingly
166 viable option. As demonstrated by the return of algae blooms to Lake Erie ^{24, 30}, P dynamics are
167 complex, requiring vigilance to incorporate both new and historical information into adaptive
168 management. Improved understanding of long-term time lags for transport ¹⁰, and more timely
169 updates to spatially- and temporally-explicit data sets on traded goods and wastes containing P,
170 may help identify strategies that sustain food production while protecting water quality.

171

172 Methods

173 We used both published and new data on major P fluxes across the boundaries of the
174 landscape P pool (soils+aquatic systems), as well as within-basin P transfers. Methods for the net
175 annual P input calculations were informed by known properties of each basin, including
176 physiographic setting, human population, and size (Table S1). A summary of the sources of P
177 flux data and calculations is provided in Table S2. The time series for each P flux, and net annual
178 P inputs, are provided in Table S3 (Maumee), Table S4 (Thames), and Table S5 (Yangtze), and
179 we used discrete time in annual intervals. Three linked reasons for our focus on Maumee,
180 Thames, and Yangtze Basins are: 1) each basin has major human influences that may relate to
181 the long-term P dynamics; 2) there have been major management, monitoring, and research
182 efforts in these basins for several decades, leading to the P data sets that provide a unique
183 opportunity to reconstruct the long-term net P inputs to soils and aquatic systems; 3) the basins

184 differ substantially in terms of socio-economic history and physiographic features but are linked
185 by common interests of water security, food security, and resource management.

186 We define the basin-level net annual P input (P_{net} , mass per year) as

$$187 \quad P_{net} = P_{in} - P_{out} \quad (1)$$

188 where P_{in} is gross annual input and P_{out} is gross annual output to/from the landscape P pool. In
189 our conceptualization, human systems such as markets, waste treatment facilities, and landfills
190 are not components of the landscape P pool, but still may greatly influence it through exchange.
191 Note that the calculations of P_{net} , P_{in} , and P_{out} were not merely the summation of the simple
192 component fluxes plotted in Fig. 1, which includes internal transfers within the basin. Rather, the
193 net/gross calculations required more thorough book-keeping of new/exogenous P inputs and
194 permanent outputs across the basin boundaries, not double-counting of the same P mass moved
195 internally. Gross inputs from equation 1 may be broken down further as

$$196 \quad P_{in} = P_{fert,in} + P_{sewage,in} + P_{precip} \quad (2)$$

197 where P_{precip} is atmospheric P input from precipitation, $P_{fert,in}$ is gross mineral fertilizer P import
198 via trade, and $P_{sewage,in}$ is the subset of sewage P production that originates from imported
199 products (food + household cleaners) and enters the environment either as effluent from sewage
200 treatment or as biosolids/sludge waste applied to soils. The new landscape P input represented by
201 $P_{sewage,in}$ is not to be confused with total sewage P production plotted in Fig 1. Rather, total
202 sewage P production contains internally produced food P already accounted as fertilizer input.
203 P_{precip} in agricultural basins is often small relative to fertilizer use, as evidenced by Maumee
204 River Basin, where P_{precip} was reported to be 0.2 kt per yr²⁵, or <1% of mean fertilizer P import
205 over our period of record. Equation 2 simplifies to

$$206 \quad P_{in} = P_{fert,in} + P_{sewage,in} \quad (3)$$

207 under the assumption of $P_{precip}=0$. The outputs may be broken down further as

$$208 \quad P_{out} = P_{food/feed,out} + P_{river} \quad (4)$$

209 $P_{food/feed,out}$ is gross P export via food/feed trade and waste transport to landfills, and P_{river} is P
210 exported via fluvial transport. Note that un-mined rock-P is not a part of the landscape pool in
211 our conceptualization, so there is no need to include an export term for fertilizer P. Substituting
212 equations 3 and 4 into equation 1 gives

$$213 \quad P_{net} = P_{fert,in} + P_{sewage,in} - P_{food/feed,out} - P_{river} \quad (5)$$

214 and we used equation 5 as the central basis for constructing time series of net annual P input.
215 Accumulated P stores were quantified by taking the cumulative sum of the P_{net} (t) time series,
216 across years.

217

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290

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298

299 Author Contributions

300 S.M.P. led the writing of the paper, compiled the data, and analyzed the data. Key P data sets
301 were contributed by H.P.J., N.J.K.H., F.W., T.W.B., and J.S. All authors participated in the
302 interpretation of results and the writing and editing process.

303 Figure legends

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305 Figure 1. Component P fluxes used in calculating the net annual P inputs for the three river
306 basins (Maumee R. USA, Thames R. UK, Yangtze R. China).

307

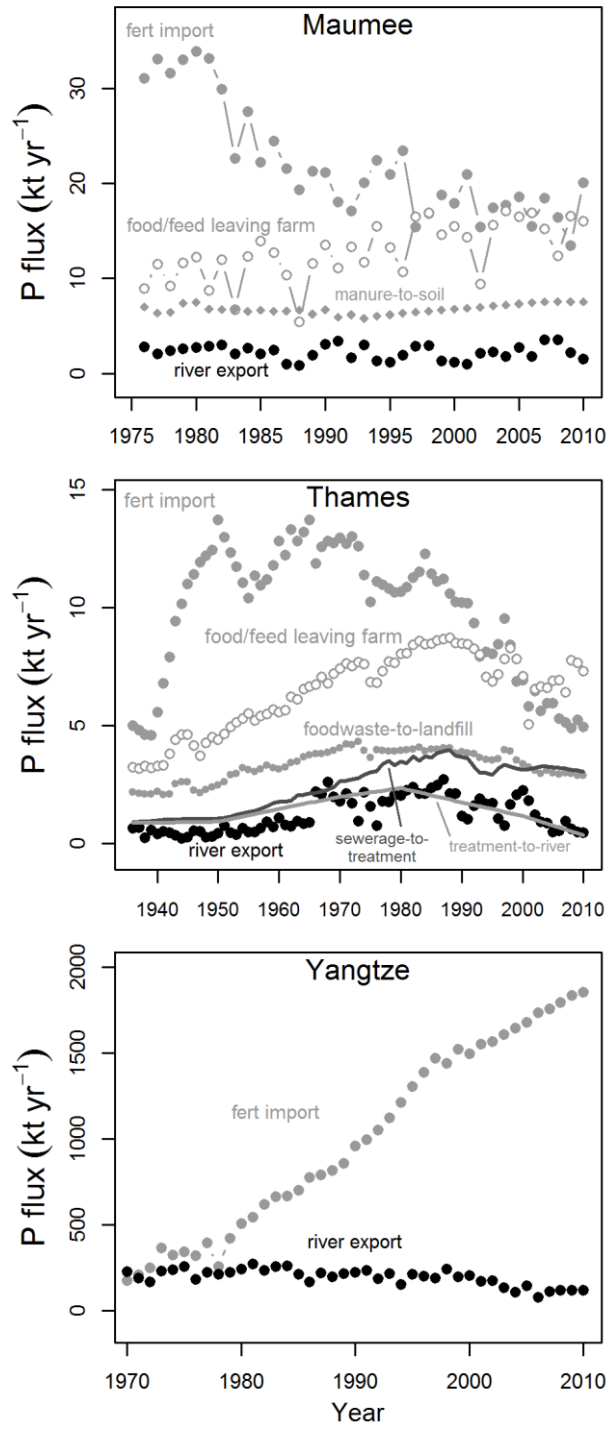
308 Figure 2. Gross P inputs and outputs to/from the landscape P pool (soils + aquatic systems) of
309 Maumee and Thames Basins. Gross P input includes fertilizer import, and for Thames only,
310 detergent import. Gross P output includes river export, food/feed exported from the basin via
311 trade, and for Thames only, disposal of foodwaste to landfill and disposal of sewage biosolids to
312 landfill, sea, or incinerator.

313

314 Figure 3. Net annual P input and accumulation curves for landscape P pools (soils+aquatic
315 systems) of three river basins (Maumee R. USA, Thames R. UK, Yangtze R. China).

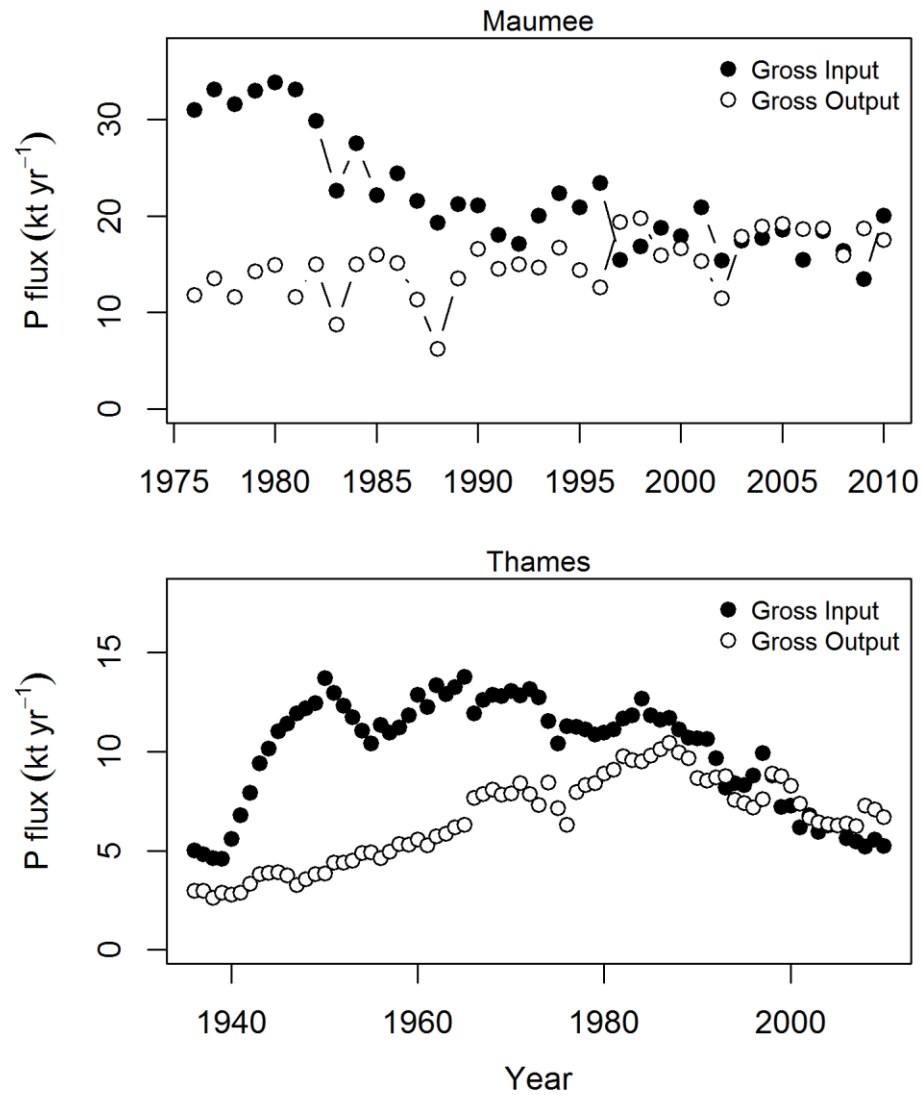
316 Accumulated P is the cumulative sum of net annual P input over time.

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321 Figure 1. Component P fluxes used in calculating the net annual P inputs for the three river
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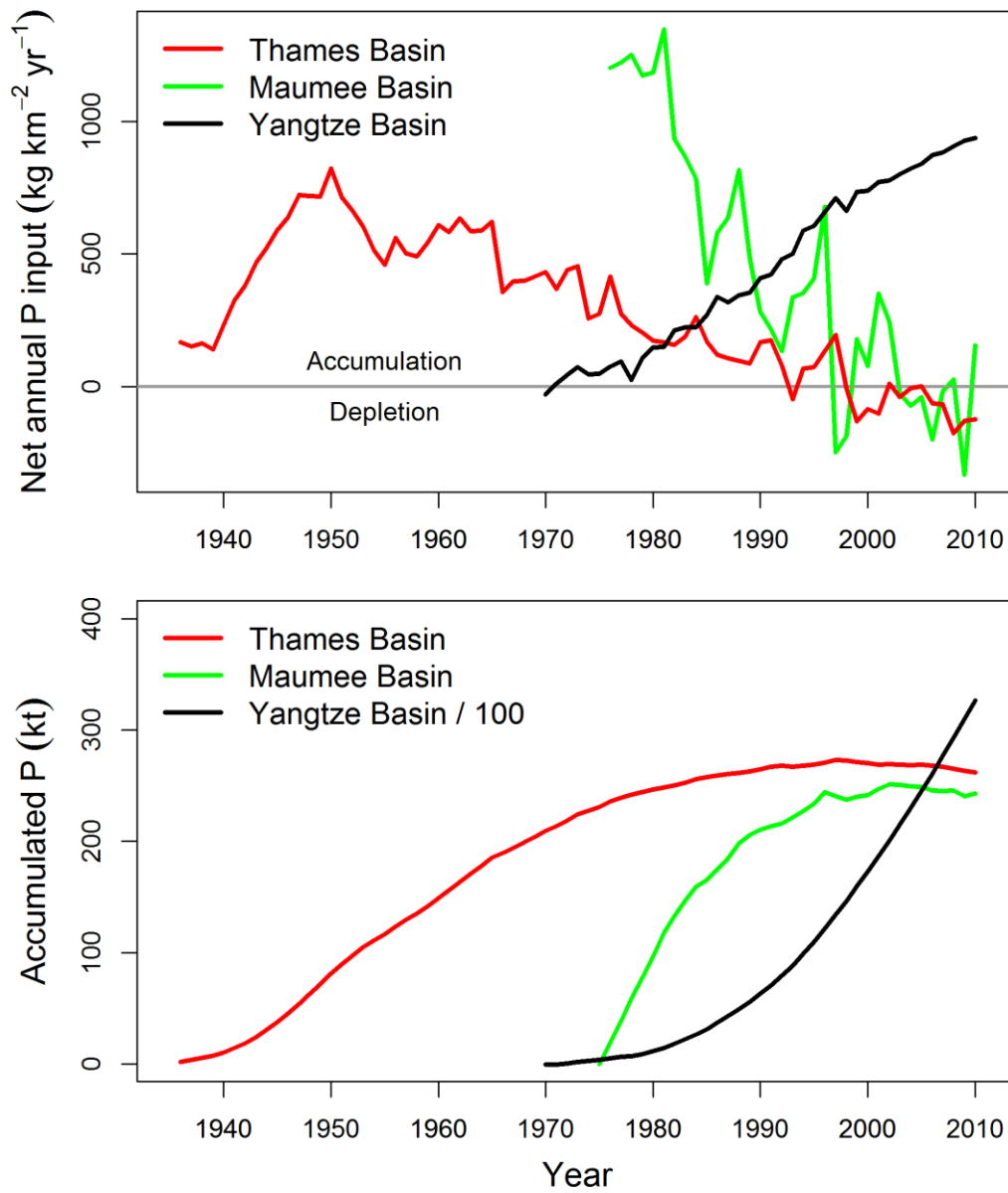
327 Maumee and Thames Basins. Gross P input includes fertilizer import, and for Thames only,

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Figure 3. Net annual P input and P accumulation curves for the landscape P pools (soils+aquatic systems) of three river basins (Maumee R. USA, Thames R. UK, Yangtze R. China). Accumulated P is the cumulative sum of net annual P input over time.